



Assessment of the influence of biophysical properties related to fuel conditions on fire severity using remote sensing techniques: a case study on a large fire in NW Spain

Journal:	<i>International Journal of Wildland Fire</i>
Manuscript ID	WF18156.R2
Manuscript Type:	Research Note
Date Submitted by the Author:	n/a
Complete List of Authors:	García-Llamas, paula; Universidad de Leon, Biodiversity and Environmental Management Suarez-Seoane, Susana; University of Leon, Area de Ecologia. Dpto. Biodiversidad y Gestion Ambiental Taboada, Angela; Universidad de Leon, Biodiversity and Environmental Management Fernández-García, Víctor; Universidad de Leon, Biodiversity and Environmental Management Fernández-Guisuraga, José Manuel; Universidad de Leon, Biodiversity and Environmental Management Fernández-Manso, Alfonso Quintano, Carmen Marcos, Elena; Universidad de Leon, Biodiversity and Environmental Management Calvo, Leonor; Univ de Leon, Fac de Ciencias Biologicas y Ambientales
Keyword:	Fire severity, Fuel, Ecosystems: temperate, Fire management

SCHOLARONE™
Manuscripts

Assessment of the influence of biophysical properties related to fuel conditions on fire severity using remote sensing techniques: a case study on a large fire in NW Spain

P. García-Llamas^{*a}, S. Suárez-Seoane, A. Taboada^a, V. Fernández-García^a, J.M. Fernández-Guisuraga^a, A. Fernández-Manso^b, C. Quintano^c, E. Marcos^a, L. Calvo^a

P. García-Llamas (corresponding author)

e-mail: pgarcl@unileon.es; phone: 0034987291567

^a Area of Ecology, Faculty of Biological and Environmental Sciences, University of León, 24071 León, Spain. ^b Agrarian Science and Engineering Department, University of León, Av. Astorga s/n, 24400 Ponferrada, Spain. ^c Electronic Technology Department, Sustainable Forest Management Research Institute, University of Valladolid, Spanish National Institute for Agriculture and Food Research and Technology (INIA), C/ Paseo del Cauce, 59, 47011 Valladolid, Spain.

Date of the manuscript draft: September 2018

Manuscript word count: 3920

1 **Abstract**

2 This study analyzes the suitability of remote sensing data from different sources
3 (Landsat 7 ETM+, MODIS and Meteosat) in evaluating the effect of fuel conditions on
4 fire severity, using a megafire (11891 ha) that occurred in a Mediterranean pine forest
5 ecosystem (NW Spain) between August 19th and 22nd, 2012. Fire severity was measured
6 via the delta Normalized Burn Ratio index. Fuel conditions were evaluated through
7 biophysical variables including: (i) the Visible Atmospherically Resistant Index and
8 mean actual evapotranspiration, as proxies of potential live fuel amount; (ii) Land
9 Surface Temperature and water deficit, as proxies of fuel moisture content.
10 Relationships between fuel conditions and fire severity were evaluated using Random
11 Forest models. Biophysical variables explained 40 % of the variance. The Visible
12 Atmospherically Resistant Index was the most important predictor, being positively
13 associated with fire severity. Evapotranspiration also positively influenced severity,
14 although its importance was conditioned by the data source. Live fuel amount, rather
15 than fuel moisture content, primarily affected fire severity. Nevertheless, an increment
16 in water deficit and land surface temperature was generally associated with greater fire
17 severity. This study highlights that fuel conditions largely determine fire severity,
18 providing useful information for defining pre-fire actions aimed at reducing fire effects.

19

20 **Keywords:** VARI index, evapotranspiration, Meteosat, MODIS, Landsat, fire effects

21

22

23

24

25

26

27 **Introduction**

28 In the European Mediterranean region, fire is a major disturbance (Oliveira *et al.* 2012)
29 with significant ecological and socio-economic impacts on forest ecosystems (Pausas *et*
30 *al.* 2009). It is well established that a major determinant of the magnitude of the ecological
31 impact and effects of wildfires is fire severity (Harris and Taylor 2017), as it can alter
32 vegetation composition, structure and regeneration dynamics (Wang and Kembell 2003;
33 González-De Vega *et al.* 2018), as well as contribute to increasing soil degradation
34 (Heydari *et al.* 2017). Fire severity refers to the change between pre- and post-fire
35 conditions (Key 2006; Meng *et al.* 2017; Fernández-García *et al.* 2018a), and is
36 operationally represented as both aboveground and belowground consumption of organic
37 matter (Keeley 2009). It has been commonly evaluated through field methods, (e.g., the
38 Composite Burn Index – CBI – and the GeoCBI index); but also using remotely sensed
39 spectral indices validated with field-measured metrics, as a timely and cost-effective
40 alternative to field methods (Fang *et al.* 2018). Properties of fire regime, such as the
41 severity and size of fires, are expected to increase in the future in the Mediterranean
42 region, likely due to land use and climate change, and forest management policies
43 (González-De Vega *et al.* 2016), which might lead to drastic shifts in fire activity and
44 seasonality. Therefore, modelling potential fire severity and understanding its main
45 drivers of control emerges as a priority for improving pre-fire forest management
46 strategies (Estes *et al.* 2017; García-Llamas *et al.* 2019).

47 Among the environmental factors that influence fire severity, there is increasing evidence
48 that fuel is a major controlling factor (Kraaij *et al.* 2018; García-Llamas *et al.* 2019). In
49 forest ecosystems, fuel characteristics, such as fuel moisture and structure, may affect fire
50 spread, progression and behaviour (Harris and Taylor 2017), which largely determines
51 fire severity levels. Furthermore, fuel composition and loading influence heat flux during

52 combustion, which ultimately may condition the spatial patterns of fire severity (Fang *et al.*
53 *et al.* 2018). Nevertheless, how fuel characteristics are specifically related to fire severity is
54 still not fully understood. Whereas studies by Lentile *et al.* (2006) and Lydersen *et al.*
55 (2017) have shown clear relationships between fuels and fire severity, others, such as
56 Bessie and Johnson (1995) and Estes *et al.* (2017), suggested that fuels have a less
57 important role on fire severity compared to other environmental factors (e.g., weather
58 conditions and topography).

59 Fuel characteristics, such as fuel amount or spatial structure, can be modified through
60 management treatments (Lee *et al.* 2018). As a consequence, knowledge of the role
61 played by fuel in fire severity is critical for prioritizing effective pre- and post-fire
62 management strategies. Fire management strategies require, however, the development
63 of reliable and accurate information that helps and supports decision-making processes
64 (Chuvieco and Kasischke 2007).

65 Recent advances in remote sensing techniques have provided major opportunities to
66 obtain valuable information for scientists and decision-makers related to fuel
67 characteristics for fire severity modelling in a cost-effective way. For example, satellite
68 remote sensing offers great potential for (i) mapping fuel models (Riaño *et al.* 2002; van
69 Wagtendonk and Root 2003); (ii) estimating live fuel moisture content from vegetation
70 indices (Myoung *et al.* 2018); and (iii) measuring potential biomass production, the
71 balance between moisture availability, fuel dryness and vegetation drought-stress from
72 remotely sensed evapotranspiration products (Kane *et al.* 2015; Fang *et al.* 2018).
73 Information from remote sensing systems offers several advantages as it is spatially
74 comprehensive and can be periodically updated (Chuvieco and Kasischke 2007), thus
75 enabling the assessment of spatial and temporal variation in fuel characteristics and their
76 effect on fire severity. For example, Landsat satellite has been widely used for monitoring

77 and modelling fuel characteristics, since it provides one of the longest moderate spatial
78 resolution imagery collections (Banskota *et al.* 2014). Moderate Resolution Imaging
79 Spectroradiometer (MODIS) vegetation products have also been commonly used in fire
80 studies across the globe, due to their near-global spatial coverage and high temporal
81 resolution (Uyeda *et al.* 2015; Fang *et al.* 2018). Additionally, characteristics of newer
82 satellites, such as the high temporal resolution of Meteosat Second Generation (MSG;
83 (Amraoui *et al.* 2013), are incurring interest in the fire research field. Nevertheless,
84 despite its advantages, the operational use of remote sensing data in assessing the role of
85 fuels in fire severity still presents some challenges associated with the current status of
86 satellite sensor technology (Chuvieco and Kasischke 2007) and the availability of the
87 spectral, spatial or temporal resolution required for operational performance (Meng and
88 Zhao 2017).

89 In this study, we aim to examine the suitability of different remote sensing sources
90 (Landsat 7 ETM+, MODIS and Meteosat) to evaluate how biophysical properties are
91 related to fuel conditions and how they can predict fire severity. Further, we provide
92 recommendations at management level for defining actions to reduce fire effects. As a
93 case study, we used a megafire that occurred in 2012 in NW Spain, which affected 11891
94 ha of a Mediterranean ecosystem dominated by *Pinus pinaster* Aiton.

95

96 **Methods**

97

98 ***Study site***

99 This study was conducted in the Sierra del Teleno mountain range (NW Spain; Fig. 1)
100 where 11891 ha burned in August, 2012 (between 19th and 22nd). The orography is

101 heterogeneous with altitude ranging from 2188 to 840 m.a.s.l. and 10% average slope.
102 Soils are acidic originated over siliceous lithology (i.e., quartzite, conglomerate,
103 sandstone and slate) with low organic matter content (Fernández-García *et al.* 2018b).
104 The climate in this area is Mediterranean. Mean annual temperature is 10 °C, with 2-3
105 months of drought in summer and a mean annual precipitation rate of 650 to 900 mm
106 (20 years averaged values covering period 1950-1999; Ninyerola *et al.* 2005). During
107 the week preceding the fire and during the fire itself, there was a heatwave that
108 increased the fire risk (Quintano *et al.* 2015). The Sierra del Teleno mountain range has
109 frequently been affected by wildfires mainly associated to dry spring-summer lightning
110 storms and anthropic causes (Santamaría 2015). Small fires have commonly burned the
111 area during winter, spring and autumn, while large fires mainly occur during the
112 summer season (July-September; Santamaría 2015). The area affected by the fire was
113 dominated by a mature natural maritime pine (*Pinus pinaster* Ait.) forest, with a tree
114 density in mature stands of 765 plants ha⁻¹. The shrubby understory community is
115 mostly dominated by *Erica australis* L. and *Pterospartum tridentatum* (L.) Willk.
116 Maritime pine populations in this area are highly adapted to intense crown fires with
117 more than 95% of mature trees bearing serotinous cones (Tapias *et al.* 2004).
118 Nevertheless, short fire return intervals (the average fire free interval has been estimated
119 at 15 years) might prevent *P. pinaster* from reaching reproductive maturity, thus
120 undermining population resilience (Taboada *et al.* 2018). The fire under consideration
121 was an extreme convective-crown-fire that completely destroyed the understory and
122 consumed the majority of tree crowns (40% of the surface burned at high severity
123 levels; Quintano *et al.* 2015). Such extreme fire severity characteristics justified the
124 selection of this fire event as a case study.

125

126 ***Fire severity***

127 Fire severity data were estimated from two Landsat 7 ETM+ images obtained on
128 September 20th, 2011 (pre-fire image) and September 20th, 2012 (post-fire image) from
129 the United States Geological Survey (USGS) Earth Explorer server
130 (<http://earthexplorer.usgs.gov/>). Image selection was conducted considering the
131 availability of cloud-free images closest to the date of the fire, aiming to avoid
132 phenological changes in the vegetation (Lecina-Díaz *et al.* 2014). We applied the
133 FLAASH algorithm (Berk *et al.* 1999; Matthew *et al.* 2003) to conduct atmospheric
134 correction of the images, which enabled us to obtain a Bottom of Atmosphere (BOA)
135 reflectance product.
136 Fire severity was calculated via the delta Normalized Burn Ratio (dNBR; Key and
137 Benson 2006; Eq. 1), an index widely used for estimating fire severity in forest systems
138 (Soverel *et al.* 2010; Whitman *et al.* 2018).

$$140 \quad dNBR = \frac{NIR - SWIR}{NIR + SWIR} \quad (\text{Eq. 1})$$

141
142 where the Near-Infrared (NIR) and the Short Wave Infrared (SWIR) bands used for
143 calculation were the NIR (B4) and the SWIR-2 (B7) bands of Landsat 7 ETM +. dNBR
144 values in unburned areas were normalized to zero by subtracting the average dNBR in
145 unburned areas outside the fire from those within the fire perimeter to account for inter-
146 annual phenological differences between pre- and post-fire images (Miller *et al.* 2009).
147 dNBR values were validated using the CBI index, which was estimated three months
148 after fire following the protocol described by Fernández-García *et al.* (2018a), which is
149 a modification of the CBI protocol developed by Key and Benson (2006). CBI values
150 ranged between 0 (unburned) and 3 (high severity) according to the burn severity scale

151 by Key and Benson (2006). They were obtained averaging the scores assigned to several
 152 variables of five vertical strata, in 54 plots of 30 m x 30 m randomly distributed across
 153 the study area. The correlation value between the spectral index and CBI was 0.88. See
 154 Fernández-García *et al.* (2018a) for further details on the dNBR validation.
 155 In this study, we used continuous dNBR values as the response variable in further
 156 analysis. Nevertheless, for easier interpretation, we also show dNBR as classified fire
 157 severity using breakpoints defined based on the CBI values: low severity, $45.898 \geq$
 158 $dNBR < 413.185$; moderate severity, $413.185 \geq dNBR < 732.565$; high severity, \geq
 159 732.565 ; by Fernández-García *et al.* (2018b) (Fig. 1).

160

161 ***Biophysical properties related to fuel conditions***

162 The biophysical properties related to fuel conditions were characterized by including
 163 metrics related to fuel loads and moisture content. We estimated the potential live fuel
 164 amount on the basis of two variables: (i) the Visible Atmospherically Resistant Index
 165 (VARI), and (ii) the mean actual evapotranspiration (AET). The VARI is an index
 166 based on the red, green and blue visible bands (Eq. 2; Gitelson *et al.* 2002), which is
 167 related to the live vegetation fraction and net primary production (Gitelson *et al.* 2002;
 168 Maguigan *et al.* 2016). It was derived from a Landsat 7 ETM+ image (30 m spatial
 169 resolution) obtained on September 20th, 2011 (the pre-fire image applied for calculating
 170 fire severity; see section 2.2 for further details on image pre-processing).

$$171 \quad VARI = \frac{R_{green} - R_{red}}{R_{green} + R_{red} - R_{blue}} \quad (\text{Eq. 2})$$

172 where R_{band} , $\text{band}=\text{green, red and blue}$ is the BOA reflectance for each band, respectively.

173 AET is related to potential biomass production and thus, to fuel amount (Kane *et al.*
 174 2015). It was calculated by averaging information acquired between June and August,
 175 2012 from two different remote sensing data sources: (i) a MSG (Schmetz *et al.* 2002;

176 Romaguera *et al.* 2012) evapotranspiration product at 10-day intervals and 3 Km spatial
177 resolution, provided by the EARS enterprise; (ii) the MOD16A2 global
178 evapotranspiration product at 8-day intervals and 500 m spatial resolution from MODIS
179 (<https://modis.gsfc.nasa.gov/data/dataproduct/mod16.php>; Hantson *et al.* 2015). We
180 selected summer months because it is the season when large fires mainly occurred in the
181 area (Santamaría 2015), and it is well established that a main factor of fire ignition and
182 propagation is the presence of fuel ready for burning (Gouveia *et al.* 2012; Russo *et al.*
183 2017), especially in crown convective fires.

184 Variables accounting for fuel moisture content included the Land Surface Temperature
185 (LST) and water deficit, which were derived from the MODIS satellite. We estimated
186 these variables for the week prior to the fire because both the high temperatures and the
187 low relative humidity of the heatwave episode during the week preceding the fire likely
188 exacerbated the effects of summer drought and, thus, fuel desiccation and flammability
189 (van Mantgem *et al.* 2013). The LST, which is expected to increase in drier vegetation
190 (Dasgupta *et al.* 2005), was computed by averaging daily information from the MODIS
191 1 Km LST product. Water deficit, at 500 m spatial resolution, was estimated as the
192 difference between PET and the AET (Kane *et al.* 2015). PET and AET were obtained
193 from the MOD16A2 global evapotranspiration product at 8-day intervals.

194

195 ***Statistical analysis***

196 In order to explore the relationship between the response variable (fire severity) and the
197 predictors (biophysical variables related to fuel conditions), we applied the Random
198 Forest (RF) machine learning algorithm (Breiman 2001), using the ‘randomForest’
199 package (Liaw and Wiener 2002) for R (R Core Team 2017) and a random sampling set
200 of 1000 pixels (1 % of pixels from the image) to build the models.

201 To avoid multicollinearity problems among the predictors, we previously checked
202 Pearson's bivariate correlations, the reached correlation values being lower than 0.60
203 (Supplementary material, Table 1).
204 The predictive power of RF was estimated through the internal out-of-bag error rates
205 (Kane *et al.* 2015). Furthermore, in order to obtain stable results, the parameter of *n*tree
206 (i.e., the number of trees to run) was set to 500 and the *m*try parameter (i.e., the number
207 of input predictors tested at each split) was established through initial tuning
208 experiments. The decrease in the accuracy (% IncMSE) criterion was used to determine
209 the relative importance of predictors in the variance explained in models. RF models
210 were run 50 times and the average was provided as the final result, aiming to obtain
211 stable model outputs and to minimize stochastic errors. Additionally, we obtained
212 partial dependence plots for each predictor.

213

214 **Results**

215 Random Forest models accounted for approximately 40% of the fire severity variance.
216 Regarding the individual contribution of each predictor in explaining fire severity,
217 biophysical properties associated with the potential amount of live fuel were relatively
218 more important than those associated with fuel moisture content (Fig. 2). In detail, the
219 VARI index emerged as the most important predictor influencing fire severity (Fig. 2).
220 Overall, high values of the VARI index were related to an increment in fire severity levels,
221 thus indicating higher fire severity in areas of great availability of live fuel (Fig. 3 a).
222 Additionally, the importance of AET in Random Forest models changed between remote
223 sensing data sources of different spatial resolution (Fig. 2). Particularly, AET obtained
224 from MSG was the second most influential predictor explaining fire severity.
225 Nevertheless, AET derived from MODIS had less influence on fire severity, even less

226 than biophysical properties related to fuel moisture content (i.e., water deficit) (Fig. 2).
227 Regardless of the remote sensing data source, higher AET values were correlated with
228 higher fire severity levels, but just towards a threshold (2.5 mm and 2.9 mm for AET from
229 MSG and MODIS, respectively; Fig. 3 b, d). Increasing water deficit was generally
230 associated with greater fire severity levels (Fig. 3 c). Furthermore, LST was weakly
231 related to fire severity (Fig. 2) and exhibited a negative influence on fire severity (Fig. 3
232 e).

233

234 **Discussion**

235

236 ***Influence of fuel on fire severity***

237 The results of this study confirm previous findings demonstrating the role of fuel
238 conditions, obtained from different remote sensing data sources, as major
239 controlling factors of fire severity patterns (Lentile *et al.* 2006; Gouveia *et al.*
240 2012; Kraaij *et al.* 2018). Nevertheless, in Mediterranean pine forest dominated
241 by *P. pinaster*, results showed that fuel characteristics were not equally related to
242 fire severity. The amount of live fuel, measured through the VARI index,
243 appeared to be the most important factor, positively affecting fire severity.
244 Positive correlations between higher levels of fire severity and the presence of
245 dense live vegetation loads has also been reported in other areas dominated by
246 pine forests (Schoennagel *et al.* 2004; Arkle *et al.* 2012). In this context,
247 chemical properties of *P. pinaster*, such as high resin content, together with the
248 structural characteristics of needles, tend to increase live biomass flammability
249 and the energy released during combustion (Calvo *et al.* 2003), therefore
250 contributing to higher fire severity levels. Additionally, recurrent fires in some

251 zones of the study site have contributed to high post-fire regeneration stand
252 densities (Calvo *et al.* 2013; Taboada *et al.* 2017), and resprouter shrub species
253 [i.e., *Erica australis* L. and *Pterospartum tridentatum* (L.) Willk.] of high
254 pyrogenicity (Calvo *et al.* 2008), which have been found to trigger high fire
255 severity levels (García-Llamas *et al.* 2019).

256 The importance of live fuel on fire severity was also evinced by the overall
257 positive effect of AET on fire severity, likely due to the association of this
258 parameter with vegetation productivity and, thus, with mounts of live fuel (Kane
259 *et al.* 2015). Nevertheless, the impact of AET on fire severity changed
260 substantially depending on the remote sensing data source used for analyses.

261 AET obtained from MSG was the second most important predictor of fire
262 severity, but the AET product from MODIS showed less importance than fuel
263 moisture predictors (i.e., water deficit). The difference in spatial resolution
264 between remote sensing derived AET products might justify this inconsistency
265 in AET importance, thus indicating that the resolution might affect the
266 predictability of fire severity models (Harris and Taylor 2017; Fang *et al.* 2018).

267 In this context, it is well known that different spatial processes could operate at
268 different scales and, hence, conclusions at one scale might not be enforceable at
269 another (Suárez-Seoane and Baudry 2002; Wu and Li 2009). Consequently,
270 spatial resolution discrepancies between data sources may constrain the accuracy
271 of models and lead to conflicting conclusions, thus limiting the development of
272 remote sensing applications (Wu and Li 2009; García-Llamas *et al.* 2016). As a
273 result, although the capacity of remote sensing techniques to provide information
274 at multiple resolutions might be advantageous (Lentile *et al.* 2006), their utility
275 for assessing the role of fuel on fire severity might be hampered by mismatches

276 between the resolution of the data source and the scale at which fuel
277 characteristics and fire severity correlate.

278 High fire severity levels have proven to be largely determined by fuel moisture
279 content (Ferguson *et al.* 2002). Our results indicated that high-severity fires were
280 more likely under greater hydric stress conditions (i.e., higher water deficit and
281 LST values). This result might be explained by the fact that dry conditions tend
282 to favour the consumption of greater amounts of fuel, as well as higher levels of
283 energy released during combustion (Dillon *et al.* 2011). Nevertheless, although
284 summers in the Mediterranean Iberian Peninsula are typically dry enough to
285 promote fuel desiccation that permits ignition, the abundance of live biomass
286 loads for combustion, rather than fuel moisture, has been noted as the primary
287 limiting factor of fire severity (Pausas and Paula 2009; Lecina-Diaz *et al.* 2014),
288 as also observed in our study. One reason could be that dry conditions limit
289 vegetation growth and, thus, fuel accumulation and continuity, leading to a
290 decrease in the risk of crown fire spread (Alvarez *et al.* 2012) and fire severity.
291 Additionally, these results could also be related to scale issues, in a way that the
292 spatial resolution of moisture predictors may not properly match the scale at
293 which fire severity patterns and fuel moisture content characteristics correlate.

294

295 ***Management recommendations***

296 Our findings evinced how high live fuel accumulations may increase
297 susceptibility to high-severity fire events in Mediterranean *P. pinaster* forest
298 ecosystems. Under this assumption, pre-fire management strategies aiming at
299 reducing high live fuel loads would be essential to reduce the likelihood of
300 severe fires. Effective pre-fire fuel treatments should prioritize the reduction of

301 canopy bulk density through silvicultural treatments, aiming at hampering crown
302 fire spread, and dismissing fire intensity, as well as convective heat transfer into
303 the canopy, thus reducing fire severity (Lininger 2006). Additionally, creating
304 open and sparse stands and retaining large trees, which reduce fuel continuity,
305 would also be recommended, aiming to increase the resilience of the system
306 (Agee and Skinner 2005). In this way, studies by Gallegos *et al.* (2003) and Kim
307 *et al.* (2016) showed how a relatively open forest structure was correlated with a
308 decrease in fire severity. Nevertheless, it is necessary to consider that fuel
309 reduction treatments need to be balanced against the development of fire-prone
310 understory vegetation. In this context, stand opening might enhance the
311 development of fire-prone shrubby understory (Fernandes and Rigolot 2007) and
312 the desiccation of live and dead fuels (Peterson *et al.* 2003), which would make
313 periodic surface fuel treatments necessary.

314

315 **Conclusions**

316 The results of this study highlight that, in severe crown-convective fires in *P. pinaster*
317 Mediterranean forest, the accumulation of live vegetation available to be burned plays a
318 relatively more important role in determining high levels of fire severity than fuel
319 moisture conditions. In addressing the role of fuel characteristics in fire severity, the
320 VARI index from Landsat 7 ETM+ and the AET product from MSG might be valuable
321 tools for determining the amount of live fuel susceptible to influencing fire severity.
322 However, we further highlight the importance of a proper selection of the remote data
323 sources at the operational spatial resolution which might affect the predictability of fire
324 severity models. Our analysis provides information that can be helpful for
325 environmental managers when defining strategies aimed at reducing severity and its

326 ecological effects during the pre- and post-fire decision-making process. These
327 strategies should prioritize the reduction of live fuel accumulations and the
328 enhancement of a more open canopy through the modification of forest stands and
329 structure.

330

331 **Conflicts of interest**

332 The authors declare no conflicts of interest.

333

334 **Acknowledgements**

335 Financing for this study was provided by the Spanish Ministry of Economy and
336 Competitiveness (GESFIRE project, AGL2013-48189-C2-1-R; FIRESEVES project,
337 AGL2017-86075-C2-1-R), the Regional Government of Castile and León (FIRECYL
338 project, LE033U14; SEFIRECYL project, LE001P17), and the European Regional
339 Development Fund. Further, V. Fernández-García and J.M. Fernández-Guisuraga were
340 supported by a predoctoral fellowship from the Ministry of Education, Culture and
341 Sport of Spain (FPU14/00636 and FPU16/03070).

342

343 **References**

- 344 Alvarez A, Gracia M, Vayreda J, Retana J (2012) Patterns of fuel types and Crown fire
345 potential in *Pinus halepensis* forests in the Western Mediterranean Basin. *Forest*
346 *Ecology and Management* **270**, 282-290.
- 347 Agee JK, Skinner CN (2005) Basic principles of forest fuel reduction treatments. *Forest*
348 *Ecology and Management* **211**, 83-96.
- 349 Amraoui M, Liberato MLR, Calado TJ, DaCamara CC, Pinto-Coelho L, Trigo RM,
350 Gouveia CM (2013) Fire activity over Mediterranean Europe based on information
351 from Meteosat-8. *Forest Ecology and Management* **294**, 62-75.
- 352 Arkle RS, Pilliod DS, Welty JL (2012) Pattern and process of prescribed fires influence
353 effectiveness at reducing wildfire severity in dry coniferous forests. *Forest Ecology*
354 *and Management* **276**, 174-184.
- 355 Banskota A, Kayastha N, Falkowski MJ, Wulder MA, Froese RE, White JC (2014) Forest
356 monitoring using Landsat time series data: a review. *Canadian Journal of Remote*
357 *Sensing* **40**, 362-384.
- 358 Berk A, Anderson GP, Bernstein LS, Archarya PK, Dothe H, Matthew MW, Michael W,
359 Adler-Golden SM, Chetwynd JH, Richtsmeier SC, Pukall B, Allred CL, Jeong LS,
360 Hoke ML (1999) MODTRAN4 radiative transfer modelling for atmospheric
361 correction. *Proceedings of SPIE* **3756**, 348-353.
- 362 Bessie WC, Johnson EA (1995) The relative importance of fuels and weather on fire
363 behavior in subalpine forests. *Ecology* **76(3)**, 747-762.
- 364 Breiman L (2001) Random forests. *Machine learning* **45(1)**, 5-32.
- 365 Calvo L, Santalla S, Marcos E, Valbuena L, Tárrega R, Luis E (2003) Regeneration after
366 wildfire in communities dominated by *Pinus pinaster*, an obligate seeder, and in others

367 dominated by *Quercus pyrenaica*, a typical resprouter. *Forest Ecology and*
368 *Management* **184**, 209-223.

369 Calvo L, Santalla S, Valbuena L, Marcos E, Tárrega R, Luis-Calabuig E (2008) Post-
370 fire natural regeneration of a *Pinus pinaster* forest in NW Spain. *Plant Ecology* **197**,
371 81-90.

372 Calvo L, Santalla S, Valbuena L, Marcos E, Tárrega R, Luis-Calabuig E (2008) Post-
373 fire natural regeneration of a *Pinus pinaster* forest in NW Spain. *Plant Ecology* **197**,
374 81-90.

375 Chuvieco E, Kasischke ES (2007) Remote sensing information for fire management and
376 fire effects assessment. *Journal of Geophysical Research* **112**, G01S90.

377 Dasgupta S, Qu JJ, Hao X (2005) Evaluating remotely sensed live fuel moisture
378 estimations for fire behavior predictions. EastFIRE Conference. (Fairfax, VA).

379 Dillon GK, Holden ZA, Morgan P, Crimmins MA, Heyerdahl EK, Luce CH (2011) Both
380 topography and climate affected forest and woodland burn severity in two regions of
381 the western US, 1984 to 2006. *Ecosphere* **2(12)**, 1-33.

382 Estes BL, Knapp EE, Skinner CN, Miller JD, Preisler HK (2017) Factors influencing
383 fire severity under moderate burning conditions in the Klamath Mountains, northern
384 California, USA. *Ecosphere* **8(5)**, e01794.

385 Fang L, Yang J, White M, Liu Z (2018) Predicting potential fire severity using
386 vegetation, topography and surface moisture availability in a Eurasian boreal forest
387 landscape. *Forests* **9(3)**, 130.

388 Ferguson SA, Ruthford JE, McKay SJ, Wright D, Wright C, Ottmar R (2002)
389 Measuring moisture dynamics to predict fire severity in longleaf pine forests.
390 *International Journal of Wildland Fire* **11(4)**, 267-279.

- 391 Fernandes PM, Rigolot E (2007) The fire ecology and management of maritime pine
392 (*Pinus pinaster* Ait.). *Forest Ecology and Management* **241**, 1-13.
- 393 Fernández-García V, Santamarta M, Fernández-Manso A, Quintano C, Marcos E, Calvo
394 E (2018a) Burn severity metrics in fire-prone pine ecosystems along a climatic
395 gradient using Landsat imagery. *Remote Sensing of Environment* **206**, 205-217.
- 396 Fernández-García V., Quintano C, Taboada A, Marcos E, Calvo L, Fernández-Manso A
397 (2018b) Remote sensing applied to the study of fire regime attributes and their
398 influence on post-fire greenness recovery in pine ecosystems. *Remote Sensing* **10(5)**,
399 733.
- 400 Gallegos V, Navarro RM, Fernández P, Valle G (2003) Postfire regeneration in *Pinus*
401 *pinea* L. and *Pinus pinaster* Aiton in Andalusia (Spain). *Environmental Management*
402 **31(1)**, 86-99.
- 403 García-Llamas P, Calvo L, Álvarez-Martínez JM, Suárez-Seoane S (2016) Using
404 remote sensing products to classify landscape. A multi-spatial resolution approach.
405 *International Journal of Applied Earth Observation and Geoinformation* **50**, 95-105.
- 406 García-Llamas P, Suárez-Seoane S, Taboada A, Fernández-Manso A, Quintano C,
407 Fernández-García V, Fernández-Guisuraga JM, Marcos E, Calvo L (2019)
408 Environmental drivers of fire severity in extreme fire events that affect
409 Mediterranean pine forest ecosystems. *Forest Ecology and Management* **433**, 24-32.
- 410 Gitelson AA, Stark R, Grits U, Rundquist D, Kaufman Y, Derry D (2002) Vegetation
411 and soil lines in visible spectral space: a concept and technique for remote estimation
412 of vegetation fraction. *International Journal of Remote Sensing* **23(13)**, 2537–2562.
- 413 González-De Vega S, De las Heras J, Moya D (2016) Resilience of Mediterranean
414 terrestrial ecosystems and fire severity in semiarid areas: responses of Aleppo pine

- 415 forests in the short, mid and long term. *Science of the Total Environment* **573**, 1171-
416 1177.
- 417 González-De Vega S, De las Heras J, Moya D (2018) Post-fire regeneration and
418 diversity response to burn severity in *Pinus halepensis* Mill. *Forest Ecology and Management* **418**, 299.
- 419 Gouveia CM, Bastos A, Trigo RM, DaCamara CC (2012) Drought impacts on
420 vegetation in the pre- and post-fire events over Iberian Peninsula. *Natural Hazards*
421 *and Earth System Sciences* **12**, 3123-3137.
- 422 Harris L, Taylor AH (2017) Previous burns and topography limit and reinforce fire
423 severity in a large wildfire. *Ecosphere* **8(11)**, e02019.
- 424 Heydari M, Rostamy A, Najafi F, Dey DC (2017) Effect of fire severity on physical and
425 biochemical soil properties in Zagros oak (*Quercus brantii* Lindl.) forests in Iran.
426 *Journal of Forestry Research* **28(1)**, 95-104.
- 427 Kane VR, Lutz JA, Cansler CA, Povak NA, Churchill DJ, Smith DF, Kane JT, North
428 PM (2015) Water balance and topography predict fire and forest structure patterns.
429 *Forest Ecology and Management* **338**, 1-13.
- 430 Keeley JE (2009) Fire intensity, fire severity and burn severity: a brief review and
431 suggested usage. *International Journal of Wildland Fire* **18(1)**, 116-126.
- 432 Key CH (2006) Ecological and sampling constraints on defining landscape fire severity.
433 *Fire Ecology* **2(2)**, 34-59.
- 434 Key CH, Benson NC (2006) Landscape Assessment (LA) sampling and analysis methods.
435 USDA Forest Service General Technical Reports RMRS-GTR-164-CD.
- 436 Kim DW, Chung W, Lee B (2016) Exploring tree crown spacing and slope interaction
437 effects on fire behavior with a physics-based fire model. *Forest Science and*
438 *Technology* **12(4)**, 167-175.

- 439 Kraaij T, Baard JA, Arndt J, Vhengani L, van Wilgen BW (2018) An assessment of
440 climate, weather, and fuel factors influencing a large, destructive wildfire in the
441 Knysna region, South Africa. *Fire Ecology* **14(4)**, 1-12.
- 442 Lecina-Diaz J, Alvarez A, Retana J (2014) Extreme fire severity patterns in
443 topographic, convective and wind-driven historical wildfires of Mediterranean pine
444 forests. *Plos ONE* **9(1)**, e85127.
- 445 Lee HJ, Choi YE, Lee SW (2018) Complex relationships of the effects of topographic
446 characteristics and susceptible tree cover on burn severity. *Sustainability* **10(2)**, 295.
- 447 Lentile LB, Holden ZA, Smith AMS, Falkowski MJ, Hudak AT, Morgan P, Lewis SA,
448 Gessler PE, Benson NC (2006) Remote sensing techniques to assess active fire
449 characteristics and post-fire effects. *International Journal of Wildland Fire* **15(3)**,
450 319-345.
- 451 Liaw A, Wiener M (2002) Classification and regression by randomForest. *R News* **2**, 18-
452 22. <http://CRAN.R-project.org/doc/Rnews/>
- 453 Lininger JC (2006) Effectiveness of stand-scale forest restoration, Siskiyou Mountains,
454 Oregon. Thesis dissertation. (University of Montana: Montana).
- 455 Lydersen JM, Collins BM, Brooks ML, Matchett JR, Shive KL, Povak NA, Kane VR,
456 Smith DF (2017) Evidence of fuels management and fire weather influencing fire
457 severity in an extreme fire event. *Ecological Applications* **27(7)**, 2013-2030.
- 458 Maguigan M, Rodgers J, Dash P, Meng Q (2016) Assessing net primary production in
459 montane wetlands from proximal, airborne and satellite remote sensing. *Advances in*
460 *Remote Sensing* **5(2)**, 118-130.
- 461 Matthew MW, Alder-Golden SM, Berk A, Felde G, Anderson GP, Gorodetzky D,
462 Paswaters S, Shippert M (2003). Atmospheric correction of spectral imagery:

- 463 evaluation of the FLAASH algorithm with AVIRIS data. *Proceedings of SPIE* **5093**,
464 474- 482.
- 465 Meng R, Zhao F (2017) Remote sensing of fire effects. A review for recent advances in
466 burned area and burn severity mapping. In ‘Remote Sensing of Hydrometeorological
467 Hazards’. (Eds GP Petropoulos, T Islam) 525 pp. (CRC Press: Boca Ratón).
- 468 Meng R, Wu J, Schwager KL, Zhao F, Dennison PE, Cook BD, Brewster K, Green TM,
469 Serbin SP (2017) Using high spatial resolution satellite imagery to map forest burn
470 severity across spatial scales in a Pine Barrens ecosystem. *Remote Sensing of*
471 *Environment* **191**, 95-109.
- 472 Miller JD, Safford HD, Crimmins M, Thode AE (2009) Quantitative evidence for
473 increasing forest fire severity in the Sierra Nevada and southern Cascade Mountains,
474 California and Nevada, USA. *Ecosystems* **12(1)**, 16-32.
- 475 Myoung B, Kim SH, Nghiem SV, Jia S, Whitney K, Kafatos MC (2018) Estimating live
476 fuel moisture from MODIS satellite data for wildfire danger assessment in Southern
477 California USA. *Remote Sensing* **10(1)**, 87.
- 478 Ninyerola M, Pons X, Roure JM (2005) Atlas Climático Digital de la Península Ibérica.
479 Metodología y Aplicaciones en Bioclimatología y Geobotánica. (Universidad
480 Autónoma de Barcelona, Bellaterra: Barcelona).
- 481 Oliveira S, Oehler F, San-Miguel-Ayanz J, Camia A, Pereira JMC (2012) Modelling
482 spatial patterns of fire occurrence in Mediterranean Europe using Multiple
483 Regression and Random Forest. *Forest Ecology and Management* **275**, 117-129.
- 484 Pausas JG., Llovet, J., Rodrigo, A., Vallejo, R. (2008) Are wildfires a disaster in the
485 Mediterranean basin? – A review. *International Journal of Wildfire* **17**, 713-723.
- 486 Pausas JG., Paula S (2012) Fuel shapes the fire–climate relationship: evidence from
487 Mediterranean ecosystems. *Global Ecology and Biogeography* **21**, 1074-1082.

- 488 Peterson DL, Johnson MC, Agee JK, Jain TB, McKenzie D, Reinhardt ED (2003) Fuels
489 planning: managing forest structure to reduce fire hazard. Proceedings of the Second
490 International Wildland Fire Ecology and Fire Management Congress. (Orlando:
491 Florida).
- 492 Quintano C, Fernández-Manso A, Calvo L, Marcos E, Valbuena L (2015) Land surface
493 temperature as potential indicator of burn severity in forest Mediterranean
494 ecosystems. *International Journal of Applied Earth Observation and Geoinformation*
495 **36**, 1-12.
- 496 R Core Team (2017) R: a language and environment for statistical computing. Available
497 from: <https://www.R-project.org/>
- 498 Riaño D, Chuvieco E, Salas J, Palacios-Orueta A, Bastarrika A (2002) Generation of
499 fuel type maps from Landsat-TM images and auxiliary data in Mediterranean
500 ecosystem. *Canadian Journal of Forest Research* **32(8)**, 1301-1315.
- 501 Romaguera M, Krol MS, Salama MS, Hoekstra AY, Su Z (2012) Determining irrigated
502 areas and quantifying blue water use in Europe using remote sensing Meteosat
503 Second Generation (MSG) products and Global Land Data Assimilation System
504 (GLDAS) data. *Photogrammetric Engineering & Remote Sensing* **78(8)**, 861-873.
- 505 Russo A, Gouveia CM, Páscoa P, DaCamara CC, Sousa PM, Trigo RM (2017)
506 Assessing the role of drought events on wildfires in the Iberian Peninsula.
507 *Agricultural and Forest Meteorology* **237**, 50-59.
- 508 Santamaría JE (2015) El pino pinaster ne la Sierra del Teleno historia, ordenación,
509 crecimiento y producción. Thesis dissertation. (University of León: León).
- 510 Schmetz J, Pili P, Tjemkes S, Just D, Kerkmann J, Rota S, Ratier A (2002) An
511 introduction to Meteosat Second Generation (MSG). *American Meteorological*
512 *Society* **83(7)** 977-992.

- 513 Schoennagel T, Veblen TT, Romme WH (2004) The interaction of fire, fuels and
514 climate across Rocky Mountain Forests. *BioScience* **54(7)**, 661-676.
- 515 Soverel NO, Perrakis DDB, Coops NC (2010) Estimating burn severity from Landsat
516 dNBR and RdNBR indices across western Canada. *Remote Sensing of Environment*
517 **114(9)**, 1896-1909.
- 518 Suárez-Seoane S, Baudry J (2002) Scale dependence of spatial patterns and cartography
519 on the detection of landscape change: relationships with species' perception.
520 *Ecography* **25(4)**, 499-511.
- 521 Taboada A, Tarrega R, Marcos E, Valbuena L, Suárez-Seoane S, Calvo L (2017) Fire
522 recurrence and emergency post-fire management influence seedling recruitment and
523 growth by altering plant interactions in fire-prone ecosystems. *Forest Ecology and*
524 *Management* **402**, 63-75.
- 525 Taboada A, Fernández-García V, Marcos E, Calvo L (2018) Interactions between large
526 high-severity fires and salvage logging on a short return interval reduce the regrowth
527 of fire-prone serotinous forests. *Forest Ecology and Management* **414**, 54-63
- 528 Tapias R, Climent J, Pardos JA, Gil L (2004) Life histories of Mediterranean pines
529 *Plant Ecology* **171**, 53-68.
- 530 Uyeda KA, Stow DA, Riggan PJ (2015) Tracking MODIS NDVI time series to estimate
531 fuel accumulation. *Remote Sensing Letters* **6(8)**, 587-596.
- 532 van Mantgem PJ, Nesmith JCB, Keifer MB, Knapp EE, Flint A, Flint L (2013) Climatic
533 stress increases forest fire severity across the western United States. *Ecology Letters*
534 **16(9)**, 1151-1156.
- 535 van Wagtenonk JW, Root RR (2003) The use of multitemporal Landsat Normalized
536 Difference Vegetation Index (NDVI) data for mapping fuel models in Yosemite
537 National Park, USA. *International Journal of Remote Sensing* **24**, 1639-1651.

- 538 Wang GG, Kembell KJ (2003) The effect of fire severity on early development of
539 understory vegetation following a stand replacing wildfire. 5th Symposium on Fire
540 and Forest Meteorology jointly with 2nd International Wildland Fire Ecology and
541 Fire Management Congress. (Orlando, FL).
- 542 Whitman E, Parisien MA, Thompson DK, Hall RJ, Skakun RS, Flannigan MD (2018)
543 Variability and drivers of burn severity in the northwestern Canadian boreal forest.
544 *Ecosphere* **9**(2), e02128.
- 545 Wu H, Li ZL (2009) Scale issues in remote sensing: a review on analysis, processing
546 and modeling. *Sensors* **9**, 1768-1793.
- 547

548 **Figures**

549

550 **Fig. 1** Location map of the study area (Sierra del Teleno, NW Spain) including a
551 pre-fire vegetation map of the burned area produced using: a) an
552 orthophotograph (year 2011) from the Spanish National Plan for Aerial
553 Orthophotography
554 (<http://centrodedescargas.cnig.es/CentroDescargas/index.jsp#>); b) the CORINE
555 Land Cover data base available for 2012; and c) a fire severity map obtained
556 using classified dNBR values derived from Landsat 7 ETM+ post-burned
557 imagery (20th September 2012) with breakpoints defined based on the CBI
558 values: low severity, $45.898 \geq \text{dNBR} < 413.185$; moderate severity, $413.185 \geq$
559 $\text{dNBR} < 732.565$; high severity, ≥ 732.565 from Fernández-García *et al.*
560 (2018b); b)

561 **Fig. 2** Relative importance, measured as % IncMSE, of variables from Random
562 Forest models explaining fire severity. Abbreviations are Actual
563 Evapotranspiration from Meteosat Second Generation satellite (AET_{MSG}) and
564 from MODIS satellite ($\text{AET}_{\text{MODIS}}$); and Land Surface Temperature (LST).

565 **Fig. 3** Partial dependence plots showing the relationship between fire severity
566 and each of the predictors included in Random Forest models: a) VARI index; b)
567 Actual Evapotranspiration from Meteosat Second Generation satellite (AET_{MSG});
568 c) Water deficit; d) Actual Evapotranspiration from the MODIS satellite
569 ($\text{AET}_{\text{MODIS}}$); e) Land Surface Temperature (LST).

570

571

572

573

For Review Only

574 **Fig. 1**

575

57

57

57

57

58

58

58

58

58

58

58

58

58

58

59

59

59

59

59

59

59

59

59

59

60

60

60

60

60

60

606

607

608

609

610

611

612

613

614

615

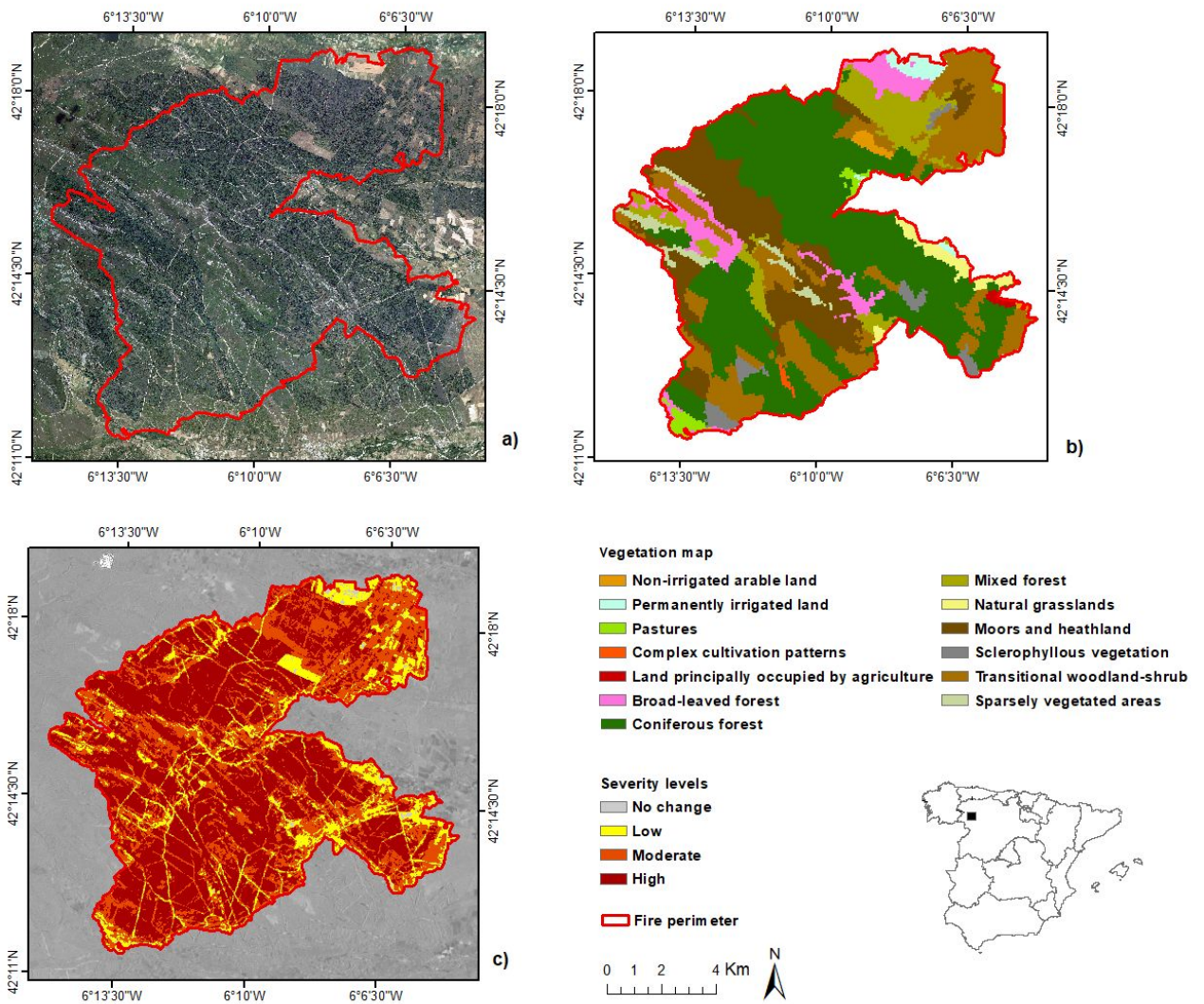
616

617

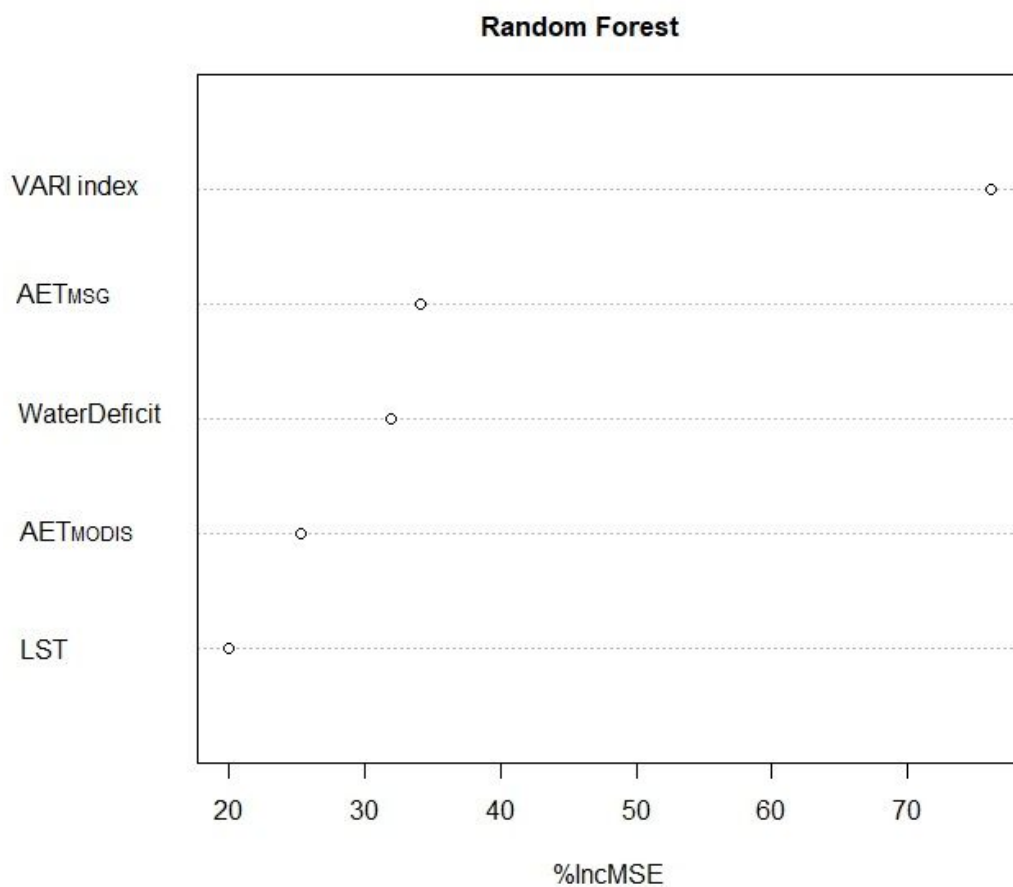
618

619

620



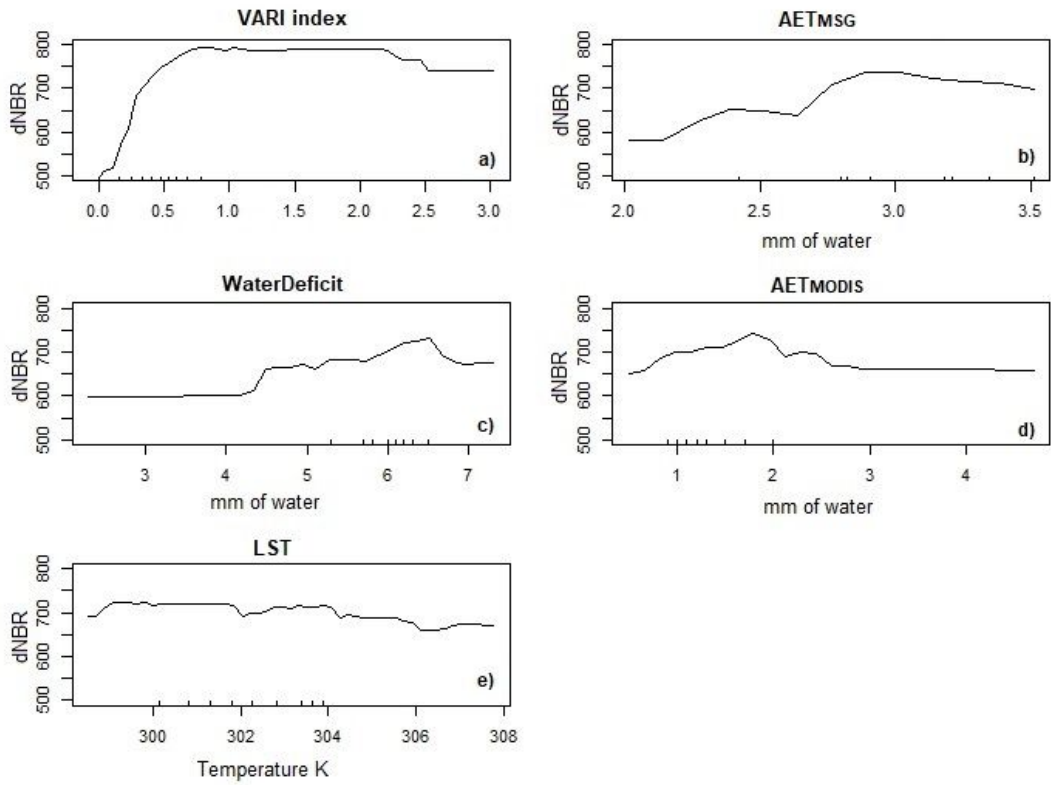
621 **Fig. 2**
 622
 623



624
 625
 626
 627
 628
 629
 630
 631
 632
 633
 634
 635
 636
 637
 638
 639
 640
 641
 642
 643
 644
 645

Only

646 **Fig. 3**
647



648

View Only

Potential live fuel amount had more influence on fire severity than fuel moisture content on pine forest ecosystems. The Visible Atmospherically Resistant Index, as a proxy of live fuel amount, showed the strongest association with fire severity. Remote sensing has high potential for determining fuel characteristics susceptible to influencing fire severity, although spatial resolution might constrain the utility of fire severity models.

For Review Only

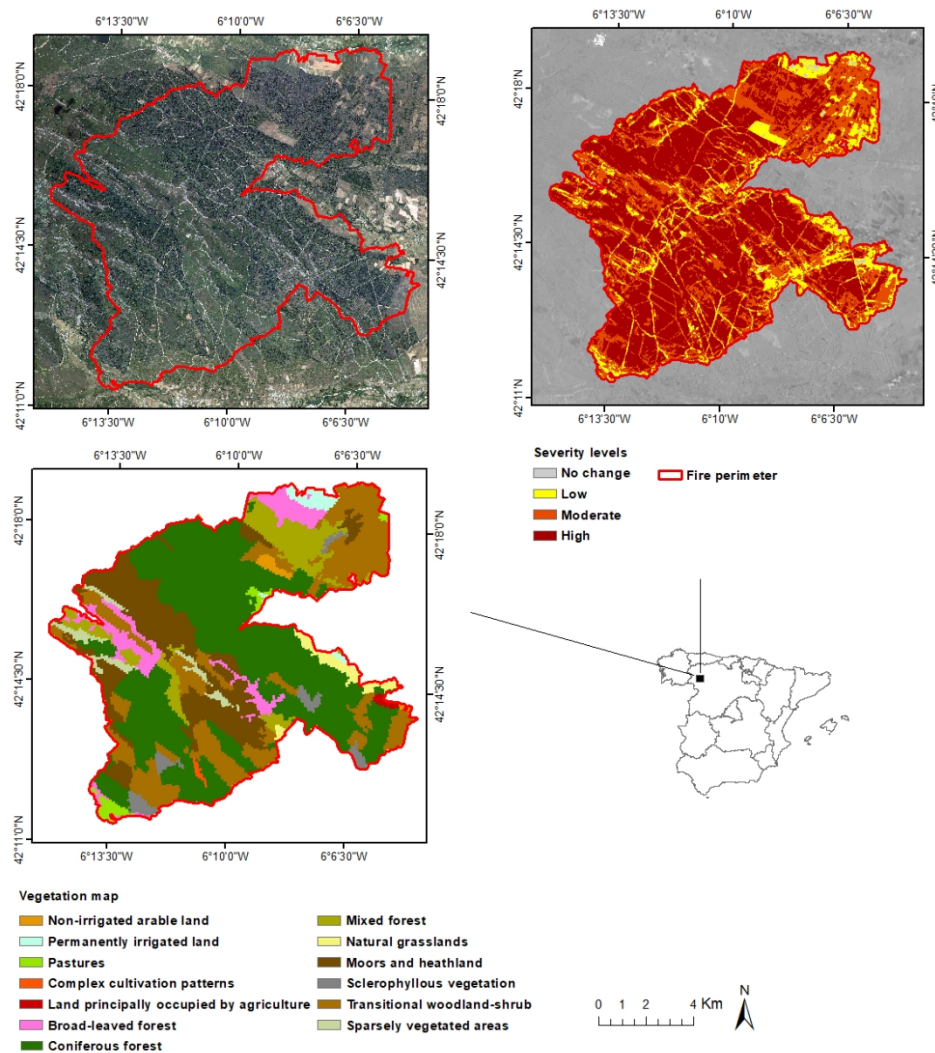


Fig. 1 Location map of the study area (Sierra del Teleno, NW Spain) including a pre-fire vegetation map of the burned area produced using: a) an orthophotograph (year 2011) from the Spanish National Plan for Aerial Orthophotography (<http://centrodedescargas.cnig.es/CentroDescargas/index.jsp#>); b) the CORINE Land Cover data base available for 2012; and c) a fire severity map obtained using classified dNBR values derived from Landsat 7 ETM+ post-burned imagery (20th September 2012) with breakpoints defined based on the CBI values: low severity, $45.898 \geq \text{dNBR} < 413.185$; moderate severity, $413.185 \geq \text{dNBR} < 732.565$; high severity, ≥ 732.565 from Fernández-García et al. (2018b); b)

312x344mm (96 x 96 DPI)

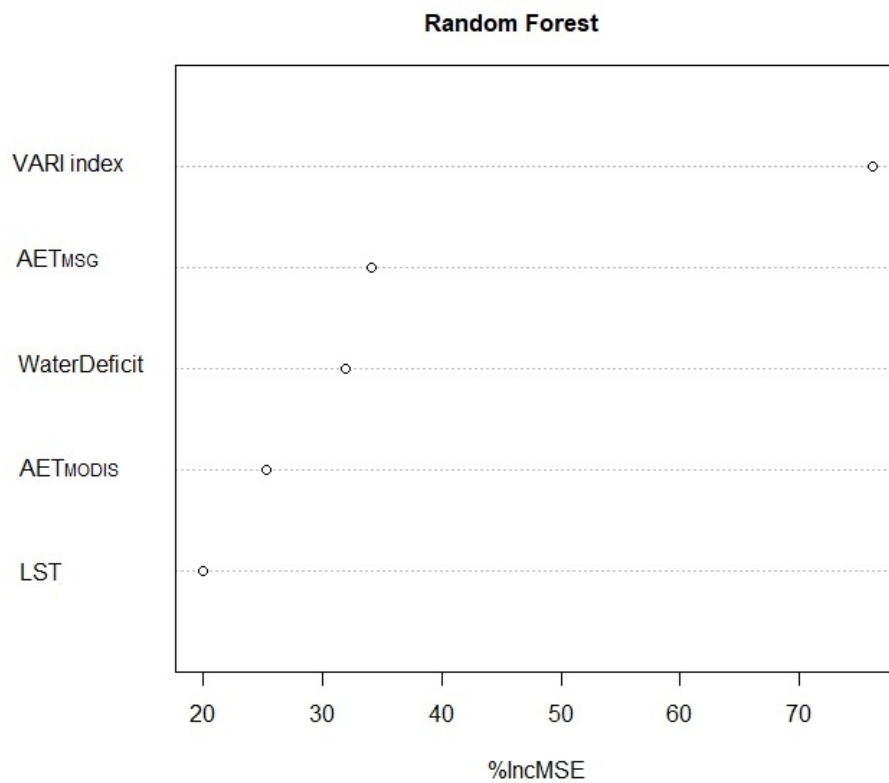


Fig. 2 Relative importance, measured as % IncMSE, of variables from Random Forest models explaining fire severity. Abbreviations are Actual Evapotranspiration from Meteosat Second Generation satellite (AETMSG) and from MODIS satellite (AETMODIS); and Land Surface Temperature (LST).

183x157mm (96 x 96 DPI)

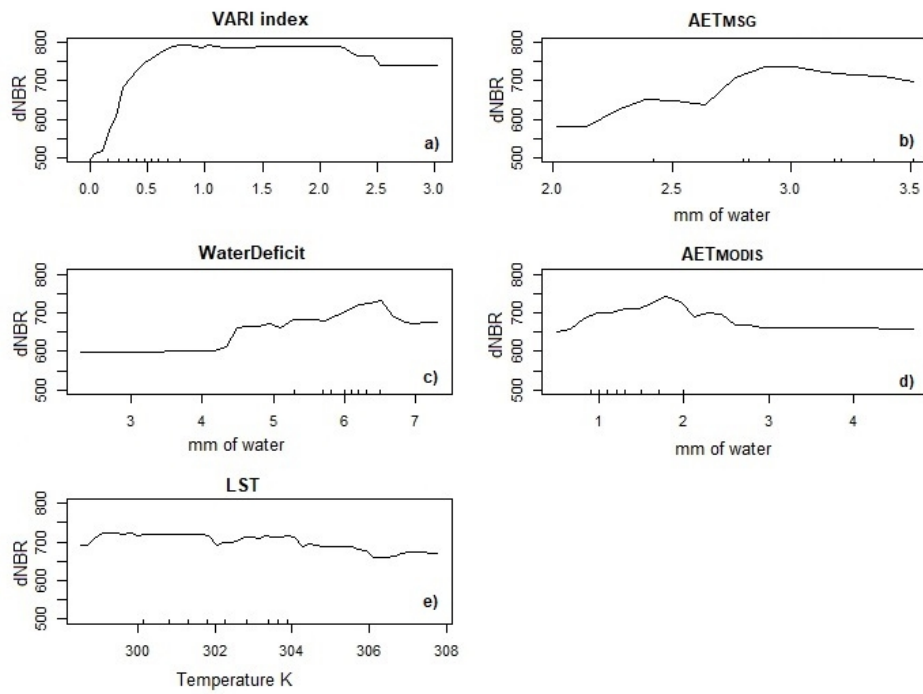


Fig. 3 Partial dependence plots showing the relationship between fire severity and each of the predictors included in Random Forest models: a) VARI index; b) Actual Evapotranspiration from Meteosat Second Generation satellite (AETMSG); c) Water deficit; d) Actual Evapotranspiration from from MODIS satellite (AETMODIS); e) Land Surface Temperature (LST).

194x139mm (96 x 96 DPI)

Table 1. Pearson's correlation coefficients (r) between pairs of predictors (biophysical variables related to fuel conditions)

	VARI index	AET _{MODIS}	AET _{MSG}	Water deficit	LST
VARI index	1.00	0.00	-0.01	0.00	-0.11
AET _{MODIS}	0.00	1.00	-0.60	0.53	-0.21
AET _{MSG}	-0.01	-0.60	1.00	-0.61	0.35
Water deficit	0.00	0.53	-0.61	1.00	-0.43
LST	-0.11	-0.21	0.35	-0.43	1.00

AET_{MODIS} (Actual Evapotranspiration obtained from the MOD16A2 global evapotranspiration product); AET_{MSG} (Actual Evapotranspiration obtained from the Meteosat Second Generation).